



Davies-Barnard, T., Valdes, P. J., Singarayer, J. S., Wiltshire, A. J., & Jones, C. D. (2015). Quantifying the relative importance of land cover change from climate and land use in the representative concentration pathways. *Global Biogeochemical Cycles*, 29(6), 842-853.
<https://doi.org/10.1002/2014GB004949>

Peer reviewed version

Link to published version (if available):
[10.1002/2014GB004949](https://doi.org/10.1002/2014GB004949)

[Link to publication record in Explore Bristol Research](#)
PDF-document

An edited version of this paper was published by AGU. Copyright 2015 American Geophysical Union. To view the published open abstract, go to <http://dx.doi.org> and enter the DOI.

University of Bristol - Explore Bristol Research

General rights

This document is made available in accordance with publisher policies. Please cite only the published version using the reference above. Full terms of use are available:
<http://www.bristol.ac.uk/red/research-policy/pure/user-guides/ebr-terms/>

1 An edited version of this paper was published by AGU. Copyright 2015 American
2 Geophysical Union.
3
4 Davies-Barnard, T., P. J. Valdes, J. S. Singarayer, A. J. Wiltshire, and C. D. Jones
5 (2015), Quantifying the relative importance of land cover change from climate and
6 land use in the representative concentration pathways, Global Biogeochem. Cycles,
7 29, 842–853. [doi:10.1002/2014GB004949](https://doi.org/10.1002/2014GB004949).
8
9 To view the published open abstract, go to <http://dx.doi.org> and enter the DOI.

10 Quantifying the relative importance of land cover change from climate and land-use
11 in the representative concentration pathways

12

13 T. Davies-Barnard^{1,2*}, P.J. Valdes¹, J.S. Singarayer³, A.J. Wiltshire⁴, C.D. Jones⁴

14

15 ¹ Cabot Institute and School of Geographical Sciences, University of Bristol, Bristol,
16 BS8 1SS, UK

17 ² College of Engineering, Mathematics and Physical Sciences, University of Exeter,
18 Laver Building, North Park Road, Exeter, EX4 4QE, UK

19 ³ Department of Meteorology, University of Reading, Reading, UK

20 ⁴ Met Office Hadley Centre, Exeter, UK

21

22 *Corresponding author address: College of Engineering, Mathematics and Physical
23 Sciences, University of Exeter, Laver Building, North Park Road, Exeter, EX4 4QE,
24 United Kingdom. Email: t.davies-barnard@exeter.ac.uk

25

Key Points

- Land area changed by climate is larger than from land-use change in the RCPs
- Climate-induced forest increases offset 90% of deforestation in RCP8.5
- Land cover change is a net carbon sink when land-use and climate are included

Abstract

Climate change is projected to cause substantial alterations in vegetation distribution, but these have been given little attention in comparison to land-use in the Representative Concentration Pathway (RCP) scenarios. Here we assess the climate-induced land cover changes (CILCC) in the RCPs, and compare them to land-use land cover change (LULCC). To do this, we use an ensemble of simulations with and without LULCC in earth system model HadGEM2-ES for RCP2.6, RCP4.5 and RCP8.5. We find that climate change causes an expansion poleward of vegetation that affects more land area than LULCC in all of the RCPs considered here. The terrestrial carbon changes from CILCC are also larger than for LULCC. When considering only forest, the LULCC is larger, but the CILCC is highly variable with the overall radiative forcing of the scenario. The CILCC forest increase compensates 90% of the global anthropogenic deforestation by 2100 in RCP8.5, but just 3% in RCP2.6. Overall, bigger land cover changes tend to originate from LULCC in the shorter term or lower radiative forcing scenarios, and from CILCC in the longer term and higher radiative forcing scenarios. The extent to which CILCC could compensate for LULCC raises difficult questions regarding global forest and biodiversity offsetting, especially at different timescales. This research shows the importance of considering the relative size of CILCC to LULCC, especially with regard to the ecological effects of the different RCPs.

54

55 Index terms: Global climate models; Earth system modelling; Land cover change;

56 Biogeochemical cycles, processes, and modelling;

57 Key Words: vegetation shifts; climate change impacts; land-use change;

58 Representative Concentration Pathways; deforestation;

59

60

61 1. Introduction

62

63 The distribution of vegetation across the globe is due to a combination of climatic and
64 anthropogenic influences, both of which are likely to alter over the next century.

65 Dynamic global vegetation models are used to project the distribution of vegetation
66 as the climate changes, and the results of this are referred to here as climate-

67 induced land cover change (CILCC). The human alterations to the land surface are
68 often known as land-use land cover change (LULCC), and encompass variations in

69 agricultural land requirement. Possible scenarios of LULCC are projected in the

70 Representative Concentration Pathways (RCPs) [Hurtt *et al.*, 2011]. The RCPs are a
71 set of future scenarios of climate change used for the 5th Climate Model Inter-

72 comparison Project (CMIP5) and the IPCC (International Panel on Climate Change)
73 5th Assessment Report [Taylor *et al.*, 2012]. They vary in their total radiative forcing

74 increase by 2100, which is indicated by the number of the RCP, (i.e. RCP8.5 has a
75 radiative forcing increase of 8.5 Watts m⁻² by 2100 compared to preindustrial levels)

76 [van Vuuren *et al.*, 2011]. The LULCC in the RCPs is prescribed by the scenario, and
77 varies over time, though it is imposed differently between models, resulting in

78 substantial variations [de Noblet-Ducoudré *et al.*, 2012]. The pattern of LULCC in the
79 RCPs has been well documented and is not linearly related to the radiative forcing of

80 the scenario [Hurtt *et al.*, 2011; Jones *et al.*, 2011; van Vuuren *et al.*, 2011; Betts *et*
81 *al.*, 2013; Brovkin *et al.*, 2013]. Notably, RCP4.5 has afforestation in the mid to high

latitudes and RCP2.6 and RCP8.5 both have tropical deforestation [Hurtt *et al.*, 2011]. LULCC in the RCPs has been extensively researched with regard to its magnitude and importance, [e.g. Thomson *et al.*, 2010; Hurtt *et al.*, 2011; Jones *et al.*, 2012; Lawrence *et al.*, 2012; Brovkin *et al.*, 2013; Davies-Barnard *et al.*, 2014a; Wilkenskjeld *et al.*, 2014]. However, changes in vegetation cover occur due to responses to climatic alterations, as well as direct human influence.

CILCC in the RCPs is simulated dynamic vegetation or some earth system models, but is not a core part of the RCP scenarios, i.e. it is a simulated response, not an imposed forcing or boundary condition. Vegetation in the models is primarily limited by temperature, water availability and carbon dioxide availability to determine the type, distribution and amount of vegetation across the globe. Very few of the CMIP5 earth system models include dynamic vegetation (that is needed to project CILCC) and therefore little work has been done on CILCC in the RCPs, especially for the time period up to 2100. Briefly discussed in the IPCC 5th Assessment Report [Ciais *et al.*, 2013], CILCC tends to be considered over longer timescales (for instance 2100 – 2300) and not in the context of LULCC. Recent research that does examine the 2005 - 2100 CILCC in the RCPs is hampered by the fact that the land cover changes are generally combined together within the standard RCP output [Betts *et al.*, 2013], making it difficult to ascertain what is LULCC and what is CILCC. Understanding CILCC is crucial to understanding both the magnitude of progressive changes (which we focus on here) but also allow the identification of potential regional ecological thresholds where abrupt and irreversible changes occur, e.g. Amazon dieback [Good *et al.*, 2012].

We aim here to highlight the importance of including CILCC in discussions of land cover change (LCC) in the RCPs. To do this, we disentangle vegetation changes induced by land-use change (LULCC), and vegetation changes induced by changes

to climate and atmospheric composition (CILCC). We use an ensemble of simulations of a selection of the RCP scenarios with and without LULCC in earth system model HadGEM2-ES (section 2). We show that for crucial aspects of environmental change in this model, such as forest and land carbon change (section 3), CILCC is often comparable and sometimes larger than LULCC. We conclude that CILCC has significant impacts for ecosystem change that are at least as big as those for LULCC (section 4) and the exact magnitude of these changes is a key research question that should be addressed.

2. Methods

2.1 Model and model simulations

We use the Met Office Hadley Centre's coupled ESM, HadGEM2-ES [Collins *et al.*, 2011; Martin *et al.*, 2011]. This coupled model includes the MOSES2 (Met Office Surface Exchange Scheme) land-surface scheme [Essery *et al.*, 2001]; the TRIFFID (Top-down Representation of Interactive Foliage and Flora Including Dynamics) dynamic global-vegetation model in dynamic mode [Cox, 2001]; the HadGEM1 physical model [Martin *et al.*, 2006]; and interactive ocean biogeochemistry, terrestrial biogeochemistry and dust and interactive atmospheric chemistry and aerosols. The atmosphere component contains 38 $1.875^\circ \times 1.25^\circ$ levels and interacts with water, energy and carbon within the land surface scheme [Essery *et al.*, 2003] and the dynamic vegetation model [Cox, 2001].

Within the dynamic vegetation land surface part of the model there are nine land surface types, including five plant functional types: broadleaf tree, needle leaf tree, C_3 and C_4 grasses and shrubs; and inland water, ice and urban. The model does not

distinguish between primary and secondary land types. The agricultural fraction is imposed as an area where broadleaf and needle leaf trees and shrubs cannot be grown. Crops are physiologically identical to grasses in the model. Increases in agricultural fraction within a gridbox are preferentially expanded into existing grass areas, only converting trees to agricultural land when the other PFTs are not available. The vegetation distribution in the model is determined by a hierarchy based on height. This results in there being a succession from grasses to shrubs and then needle leaf and broadleaf trees, as the climate becomes suitable. The dynamic global vegetation model within HadGEM2-ES, TRIFFID, is a well known and used model, extensively documented in *Cox et al.*, [1998] and *Clark et al.*, [2011]. It is one of the models used in the multi-model Global Carbon Project annual carbon budgets [Le Quéré et al., 2014a, 2014b]. It has been the land surface model for several generations of the Hadley centre climate model, and therefore used in the IPCC's assessment reports, including the most recent [Stocker et al., 2013]. The present day vegetation distribution within HadGEM2-ES is assessed in [Collins et al., 2011] and shows good agreement with present day distributions. For the tropical forests in particular, *Good et al.*, [2012] shows that the distribution in climate space validates well. The model inter-comparison by *Anav et al.*, [2013] shows that HadGEM2 has a reasonable representation of the land carbon stores.

The model setup is as for the HadGEM2-ES CMIP5 simulations [Jones et al., 2011] and the LUCID (Land-Use and Climate, IDentification of robust impacts) simulations of RCPs [Brovkin et al., 2013], using a fully dynamic atmosphere and ocean model. We use simulations of three of the RCP scenarios: RCP2.6, RCP4.5 and RCP8.5, from 2006 to 2100. Four ensemble members are initialised from historical simulations that ran from 1850 – 2005, and run for 95 years up to 2100. Two sets of simulations are used for each RCP – the standard RCP that includes LULCC, and a simulation where the agricultural fraction remains at the 2005 levels. For the simulations without

LULCC, all non land-use forcings (greenhouse gas concentrations and other aerosol forcings, etc.) are prescribed as for the equivalent RCP [Meinshausen *et al.*, 2011].

2.2 Use of Simulations

The LULCC is taken here to be the change in the agricultural fraction imposed onto the model by the RCP scenario. It is inferred from the difference between the normal 'RCP' scenarios (with LULCC) and the 'NoLUC' scenarios (without LULCC) for the last year of the simulations (2100). The CILCC is taken here to be the changes in vegetation caused by anthropogenic climate change over the period 2005 – 2100. This is inferred from the difference between the mean of the 2005 NoLUC values compared to the 2100 NoLUC values. The net changes are considered to be the standard 2005 NoLUC values compared to the RCP 2100 values. The net changes include both CILCC and LULCC changes. So the LCC calculations can be described thus:

$$\text{LULCC} = \text{RCP}^{2100} - \text{NoLUC}^{2100}$$

$$\text{CILCC} = \text{NoLUC}^{2100} - \text{Fix2005}^{2100}$$

$$\text{Net LCC} = \text{RCP}^{2100} - \text{Fix2005}^{2100}$$

Where the Fix2005 is a fixed 95 years of the 2005 land cover. Figure 1 shows how we diagnose the vegetation and carbon changes.

Even without changes in land cover, terrestrial carbon storage in biomass and soil organic matter is projected to alter due to changes in vegetation productivity, turnover, litter input to soil and soil conditions (such as temperature and moisture). Therefore to assess the CILCC separately to the accumulated vegetation carbon (not

from LCC), a control without CILCC, LULCC but with accumulated carbon is required. These were not feasible to run as fully coupled simulations due to the computational expense, so we extrapolated the control baselines of 2005 land cover including the increases to land carbon from increased carbon dioxide and temperature, but exclude the changes from LCC. These extrapolated values are used as a 'control' scenario (Fix2005) with which to infer the amount of land carbon attributable to CILCC from the anomaly. Therefore the land carbon changes can be described thus:

$$\text{LULCC carbon} = \text{RCP}^{2100} - \text{NoLUC}^{2100}$$

$$\text{CILCC carbon} = \text{NoLUC}^{2100} - \text{Fix2005}^{2100}$$

$$\text{Net LCC carbon} = \text{RCP}^{2100} - \text{Fix2005}^{2100}$$

$$\text{Accumulated carbon} = \text{RCP}^{2100} - \text{RCP}^{2006}$$

To obtain the grid box vegetation carbon, the carbon on each plant functional type (PFT) tile is weighted by the proportion of each PFT in the grid box. Therefore to approximate the vegetation carbon without any LCC, we weighted the 2100 vegetation carbon on each PFT tile by the 2005 vegetation PFT distribution (rather than the 2100 PFT distribution). This gives what the vegetation carbon would be in no LCC simulations, (excluding LULCC and CILCC, but including accumulated carbon).

To estimate the soil carbon, we take the 2005 soil carbon and scale it annually with the 2005 litter carbon and soil respiration. The soil carbon is updated each year with the input of carbon from litter carbon and then the soil respiration (which scales with the amount of soil carbon) is removed. To estimate the soil carbon, we therefore start with the 2005 soil carbon, add the litter carbon weighted by the difference between the 2005 and 'n' year PFTs, then take away the respiration weighted by proportional

difference between the 2005 soil carbon and the 'n' year soil carbon. This is repeated from n=2005 to n=2100. Thus the calculation used is:

$$CS_{nlcc}(n+1) = CS_{nlcc}(n) + [LIT_{nlcc} * (PFT_{original} / PFT_{2005}(n))] - [RH_{original}(n) * (CS_{nlcc}(n) / CS_{original}(n))]$$

where 'nlcc' is the constructed value, 'original' is the original RCP simulation value, CS is soil carbon, LIT is litter carbon, PFT is the plant functional types on tiles, and RH is soil respiration.

These offline calculations of the global soil and vegetation carbon values use the same equations as the land surface model, JULES [Clark *et al.*, 2011] that is within the coupled model. This approach has the advantage that a global value for the land carbon can be produced very efficiently and has been demonstrated as effective in other instances (for instance Liddicoat *et al.*, [2013]).

3. Results

3.1 Forest

The most notable CILCC is a global increase in forest (needle leaf and broadleaf trees) that has an approximately proportional relationship with the total radiative forcing of the scenario (see Figure 2d). This is in contrast to the LULCC, which is scenario dependent and does not have the relationship with net climate forcing that might be expected. RCP2.6 and RCP8.5 both have substantial deforestation, whereas RCP4.5 has afforestation (Figure 2d). Though RCP2.6 and RCP8.5 have very similar levels of anthropogenic deforestation, their net forest change is very

different. In RCP2.6, the CILCC offsets only 3% of anthropogenic deforestation, whereas it offsets 91% in RCP8.5. The larger increase in CILCC forest in RCP8.5 is due to higher temperature and atmospheric carbon dioxide concentration, which allows more poleward expansion of forest than in RCP2.6 (see Figure 2e, 4a and 4b).

The LULCC and CILCC forest fraction changes have noticeably different latitudinal patterns, with the tropics contributing more to LULCC and the boreal forests contributing more to CILCC. The net changes in the boreal forest latitudinal band (Figure 2e) are dominated by the CILCC increases in forest, with only relatively small LULCC. The tropics show the opposite pattern, with little CILCC and the net forest change dominated by the LULCC (see Figure 2f). Because of this, there are only a small number of isolated gridcells where both LULCC and CILCC are both strong. Globally, most of the LULCC in RCP2.6 and RCP8.5 is in the tropics, and most of the CILCC is boreal. RCP4.5 is slightly different, as there is extensive mid to high latitude afforestation due to the scenario's universal carbon tax making afforestation a viable mitigation option [Thomson *et al.*, 2010, 2011]. However, all three RCP scenarios considered here have positive net forest contributions from boreal forests, mainly from CILCC, and net tropical contributions that result mainly from LULCC.

The balance of CILCC and LULCC is different at the centennial and mid-Century time scale. The LULCC occurs relatively earlier, since LULCC agricultural expansion is instantaneous as it imposed in each year within the model. The CILCC vegetation expansion happens more gradually and therefore slightly later, as the expansion of vegetation northwards is commensurate with the increase in temperature and carbon dioxide. It also takes around 80 years in this model for abandoned agricultural land to fully reforest in the model. By 2050, globally there is very little CILCC (see Figure 2 a, b and c) and consequently there is much more influence of LULCC on the net boreal

forest LCC than at 2100. Thus the global forest amount at 2050 is more strongly influenced by the tropics and LULCC. Because of the lack of CILCC at 2050, the net LCC of RCP2.6 and RCP8.5 are much more similar than at 2100. The impact of timescale on the balance of whether LULCC or CILCC is most dominant continues further into the future. The relatively slow rate of forest growth means that for a transient climate forcing, as is projected in the RCPs, there will be committed vegetation changes for some time after the forcing stops [Jones *et al.*, 2009]. Therefore on the multi-centennial scale, CILCC is likely to be more important than LULCC.

In the tropics, there is only very slight dieback of broadleaf trees (Figure 2d and Figure 4a) in favour of C₄ grasses. Amazon dieback was a well known feature in previous versions of the Hadley Centre model (notably HadCM3) and was primarily caused by changes to precipitation over the Amazon under climate change [Cox *et al.*, 2003, 2004; Betts *et al.*, 2004; Huntingford *et al.*, 2008; Malhi *et al.*, 2009]. Amazon dieback is absent in this version of the model (HadGEM2-ES), with only up to 10% dieback over the southern edges of the Amazon (Figure 4a) [Good *et al.*, 2012]. However, since the dieback is approximately the same magnitude in all three RCPs considered here, this suggests that a relatively small change in climate may still trigger a tipping point in the Amazon in this model, which increases in carbon dioxide only very slightly compensate for (Figure 2f). In the tropics overall, this Amazon dieback is mitigated by increase in broadleaf trees over the Congo basin, where shrubs give way to broadleaf trees as the climate warms (see Figure 4a and 4c). This gives the result that in RCP2.6 the tropics has a slight decrease in forest from CILCC, but RCP4.5 has a slight increase, again aiding the mitigation of LULCC in higher radiative forcing scenarios like RCP8.5, but not RCP2.6.

3.2 All vegetation

Considering the LCC across all vegetation types, CILCC is larger than LULCC at 2100 in all the scenarios considered here (Figures 3, 4, and 5). As a per cent of global land area, CILCC is only slightly more than LULCC in RCP2.6 and RCP4.5 (CILCC: 3.2% and 5.5%; LULCC 2.9% and 5.1% respectively). However, for the high scenario, RCP8.5, the CILCC and LULCC are 8.6% and 3.9% respectively, making CILCC a factor of two bigger. The LCC values quoted above are the conservatively calculated net figures, in that no annual or decadal variations are included and the values are the simple total difference in the amount of a PFT globally between 2005 and 2100 (rather than including changes of the same land type moving to different areas) [Pongratz *et al.*, 2014; Wilkenskjeld *et al.*, 2014]. Methods of LCC calculation that included the gross changes would probably give higher CILCC values because the shifts in the PFTs would be accounted for, whereas the current method mainly accounts for the expansions. The majority of the CILCC expansion is broadleaf trees at the high latitudes (Figure 4 a) but there are shifts in vegetation all the way down the order of vegetation succession (Figures 4 and 5). As the temperature and carbon dioxide increase, more dominant or more appropriately adapted PFTs are able to move into the regions previously unable to support them. The C₃ grasses colonise furthest north, replacing the areas of bare soil and C₄ grasses (Figure 5). However, since the dynamic vegetation in the model works on a height hierarchy, shrubs and then trees have competitive advantage over grasses as the climate becomes appropriate for them, causing shrubs and then trees to move into areas previously occupied by C₃ grasses (Figures 4 and 5). Broadleaf trees are the most dominant PFT in the model, and therefore have an expansion with little dieback and the other PFTs have shifts. Thus the net change can be small even when the gross change is much more widespread, because the net change doesn't account for the shifts. Therefore the result that CILCC is larger than LULCC is likely to be robust for all the

RCP scenarios considered here, as by excluding shifts in distribution it is quite conservative.

3.3 Carbon cycle

CILCC is the largest contributor to carbon changes from net LCC and determines the signal (Figure 6a). The land carbon changes from CILCC are larger than those from LULCC in all the scenarios considered here. The net land carbon change is a sink in all three scenarios, strongly influenced by the CILCC. Soil carbon is the biggest contribution from CILCC, and is several times the size of the LULCC soil carbon change (Figure 6b). The difference in the change in soil carbon due to CILCC and LULCC is because of changes in Net Primary Production (NPP) that increase the inputs to the soil carbon [Jones and Falloon, 2009]. This is in line with the overall change in soil and vegetation carbon for all land cover (not just changed) from 2006 – 2100, which increases by 180 - 425 GtC carbon globally over the 95 year simulation (see Figure 6d, e and f). The expansion of vegetation into areas previously allocated as bare soil due to CILCC means that more litter is available to increase the soil carbon. For deforestation LULCC, the soil carbon increases a little under deforestation because some of the below ground biomass carbon goes into the soil. But the LULCC soil carbon in afforestation scenario RCP4.5 has soil carbon emissions because the trees replacing the grass have marginally lower NPP and therefore there is a loss of soil carbon. Note that the Gross Primary Production is higher for trees overall, but also trees also have higher maintenance requirements, and thus can have lower NPP. Vegetation carbon (Figure 6c) shows the opposite trend to soil carbon, with the LULCC carbon changes larger than the CILCC. The vegetation carbon changes for both CILCC and LULCC are similar to the equivalent changes in forest fraction, as in this model trees are the main stores of vegetation carbon (compare Figure 6c with Figure 2d). However, this model does not represent

any harvesting processes, which if included, would probably drive the soil carbon input down rather than up, for conversion to crops (by reducing the litter inputs when the harvest is removed elsewhere). Despite these uncertainties, these simulations suggest that net LCC is a carbon sink in all the RCPs considered here and the contribution of CILCC is larger than LULCC.

The LCC also affects the climate through changes to the atmospheric greenhouse gas concentration. The net LCC carbon change gives a cooling (Figure 6a) amounting to -0.02 K in RCP2.6, -0.21 K in RCP4.5, and -0.18 K in RCP8.5 (calculated using the HadGEM2-ES transient climate response to emissions [Gillett *et al.*, 2013]). It is notable that including CILCC changes the sign of the climate effects of net LCC in two of the RCPs. The LULCC carbon only climate impacts are +0.04 K, -0.08 K and +0.04 K (for RCP2.6, 4.5 and 8.5 respectively) [Davies-Barnard *et al.*, 2014b]. The contribution of CILCC to the carbon sink is larger than LULCC in all of the RCPs considered here, with RCP8.5 approximately four times larger. Further, the CILCC is also critical in maintaining the airborne fraction of emissions. The LULCC and increasing fossil fuel emissions historically have reduced the proportion of land-uptake of anthropogenic carbon emissions [Canadell *et al.*, 2007]. The CILCC, particularly the increase in forest fraction shown in Figure 2, means that the reduced carbon sink from LULCC is partially offset by the increase in the CILCC carbon sink [Jones *et al.*, 2012]. Therefore CILCC plays a significant role in the climatic impacts from net LCC.

4. Discussion and Conclusions

Comparing the CILCC and LULCC, we find that the CILCC has a significant impact, and in some cases a larger impact than LULCC. In all the RCPs we see a poleward

expansion and succession of vegetation, as found by field and model studies of the response of vegetation to climate changes [Emanuel *et al.*, 1985; Prentice *et al.*, 1991; Woodward *et al.*, 1998; Walther *et al.*, 2002; Soja *et al.*, 2007; Colwell *et al.*, 2008; Betts *et al.*, 2013]. The increased temperature opens up new regions that were previously too cold to support vegetation, especially in the high latitude northern hemisphere [MacDonald *et al.*, 2008]. This contrasts with LULCC in the RCPs, which is mainly in the tropics. In RCP4.5 the CILCC and LULCC globally work in parallel, giving a larger overall LCC, whereas in RCP2.6 and RCP8.5 the CILCC and LULCC offset each other.

The large CILCC in RCP8.5 means that it has a form of 'forest offsetting' over time between the deforestation in the tropics and the northward expansion of boreal forest. In RCP8.5, 91% of the anthropogenic deforestation is offset by CILCC. This could be perceived as a potential way to offset the biodiversity loss, in a similar way to biodiversity offsetting [Maron *et al.*, 2012; Reid, 2013] – compensating for the loss of tropical forest with boreal forest. However, offsetting of the total forest loss globally is an incomplete story. Tropical forests especially tend to be areas of high biodiversity [Myers *et al.*, 2000] and established primary forests are more diverse than secondary forest [Gibson *et al.*, 2011]. This could be the cause of substantial losses of global biodiversity if tropical forest were offset by boreal forest. The northward shift of forest could also cause loss of some extreme cold adapted habits. Ecosystems allocated in the model as 'bare soil' (because none of the model's plant functional types are able to sustain growth there) or C₃ grasses, could be lost entirely. It is difficult for land surface models to effectively simulate these marginal environments but they are nonetheless important and unique ecosystems.

In the short term, the net LCC would almost certainly cause losses of biodiversity. Although over the full time period to 2100 the forest changes in RCP8.5 almost

cancel out, in the period up to 2050 they do not. This question of the time lag is particular problem for biodiversity offsetting, as certain decreases are balanced against uncertain increases [Moilanen *et al.*, 2009; Bekessy *et al.*, 2010]. Probable extinctions in the tropics from LULCC would be unlikely to be meaningfully compensated for by CILCC expansion of boreal forest. Furthermore, it is possible that much of the forest gains would be not be realised, due to 'boreal dieback' from effects such as increasing destruction of forests by pests [Kurz *et al.*, 2008]. A forest offsetting policy that relied on CILCC would essentially be 'betting' on vegetation changes that may be slow or unable to be realised, whilst sacrificing established ecosystems.

From the point of view of ecosystem disruption, the greater amount of CILCC than LULCC would suggest that CILCC would cause more disruption in all three of the RCP scenarios considered here. However, habitat destruction, particularly conversion of land to agricultural use, is thought to be the most important driver of biodiversity loss, with climate change less important [Hassan *et al.*, 2005]. Since the CILCC is only slightly higher than the amount of LULCC in RCP2.6 and RCP4.5, it is possible that LULCC may have a bigger impact on biodiversity in these scenarios. For RCP8.5, CILCC would likely still be a larger impact on biodiversity, since the total area affected by CILCC is more than double than from LULCC. As well as the extent of the impact, the duration also should be taken into account. After stabilisation of the forcing, the effects of LULCC drop off, whereas the CILCC continues as the vegetation reaches equilibrium. The CILCC is likely to continue well beyond 2100 for decades or even centuries after the forcing has stabilised [Jones *et al.*, 2010; Liddicoat *et al.*, 2013]. Comparing the disruptive impact, CILCC could be a more serious challenge than LULCC, particularly in RCP8.5, because of the longevity and quantity of impact, even if the severity is lower.

The important role of CILCC in terrestrial carbon changes highlights how critical it is to reduce the uncertainty in carbon cycle projections. CILCC accounts for 14 – 22% of total terrestrial carbon changes (depending on the RCP scenario), whereas LULCC only accounts of 6 – 12% (Figure 6). Soil carbon is the biggest contributor to the land carbon change from CILCC in the model used here, around two to three times larger than vegetation carbon change. However, soil carbon change is highly variable between models, in both net sign and magnitude [Nishina *et al.*, 2014]. Some models project a global decrease in land carbon under climate change and JULES (the offline land surface model of HadGEM2-ES) is on the high side of the projections of soil carbon changes [Nishina *et al.*, 2014]. This is likely to be related to the model's sensitivity to carbon dioxide fertilisation, as this (rather than temperature) is the main driver of change in soil carbon in models [Nishina *et al.*, 2014]. Further, the vegetation carbon increase from LULCC afforestation (in RCP4.5) and CILCC may be overestimated because of lack of nitrogen limitation in the model [Gruber and Galloway, 2008; Jain *et al.*, 2013]. Conversely, the LULCC deforestation carbon change is small in HadGEM2-ES compared to other models [Brovkin *et al.*, 2013]. However, the soil carbon storage size and future sink size is highly uncertain, and its representation here is one of many possible outcomes.

The carbon effect of net LCC is also influenced by two processes not directly included in the model used in these simulations: secondary LULCC and negative emissions using bioenergy with carbon capture and storage (BECCS). The carbon changes from secondary land use changes (for instance natural to managed forest, which isn't accounted for in this model) can be substantial and may account for more carbon emissions than primary land use changes [Shevliakova *et al.*, 2009; Hurtt *et al.*, 2011; Lawrence *et al.*, 2012]. Similarly, BECCS for the RCP2.6 scenario could give negative emissions of between 43.8 to 160.6 GtC [Kato and Yamagata, 2014]. According to those projections, the potential of BECCS likely to be bigger than the

net land carbon change in any of the three RCPs considered here (8, 101 or 83 GtC for the three RCPs respectively, see figure 5 a). Therefore the lack of representation of secondary LULCC and BECCS is a considerable limitation to this study. It is also notable that the total land carbon change (including non LCC effects) is at least four times the size of the change in land carbon from LCC in this model (see figure 5 d – f). Thus the contribution of LCC to overall global carbon emissions is relatively small. However, even though the carbon effects of LCC are not substantial, other environmental impacts of LCC may be worth considering in decision making, as discussed above.

The relative lack of analysis of CILCC in the RCPs can be attributed to a combination of possible causes, including a perceived lack of need and high uncertainty. Few of the CMIP5 models include dynamic vegetation (that projects CILCC) and only around half of the CMIP5 models have vegetation carbon cycle components (19 of 38 models, [es-doc, 2014]). Although there is a slight computational cost of including dynamic vegetation to calculate CILCC in earth system models, the first implementations of the terrestrial carbon cycle were around 14 years ago [Cox *et al.*, 2000], so this is evidently not a case of inability. LULCC can be imposed onto a model using values from the Integrated Assessment Model that created the scenario, without the need for dynamic vegetation or an integrated terrestrial carbon cycle. This method excludes CILCC, and suggests a viewpoint that CILCC is not important or required. This perception is exacerbated by high uncertainty in climate-induced changes to terrestrial carbon storage. Land carbon differences within the parameter range of an individual model can be as big as the differences between the RCPs themselves [Booth *et al.*, 2012] and are highly variable between models [Nishina *et al.*, 2014]. This uncertainty presents a considerable challenge. But by neglecting to examine CILCC, we may be overestimating the importance of LULCC and misestimating land carbon change by as much as 22%.

502

503 Comparing the changes from CILCC and LULCC over 2006 – 2100, we have shown
504 that not only is the CILCC the majority of net LCC, it is also the larger part of land
505 carbon changes from net LCC. Moreover, even where CILCC is not as large as
506 LULCC, as in the case of forest change, it gives rise to issues of offsetting. To what
507 extent forest lost in the tropics could be substituted by boreal forest is both a
508 qualitative and a quantitative issue. Our results suggest that CILCC in RCP8.5 may
509 be able to quantitatively offset the deforestation, whereas it cannot in RCP2.6.
510 Whether such forest offsetting would provide equivalent ecosystem and climate
511 services is much more uncertain, and would be a useful extension to this work. Our
512 work shows that CILCC is an important aspect of the land surface in the RCPs. If the
513 potential size of the climate change impact caused or mitigated by an aspect of the
514 earth system is a guide for the amount of research that should be done on a topic,
515 then CILCC perhaps warrants more research.

516

517

518 Acknowledgements

519

520 We acknowledge funding from the Joint DECC/DEFRA Met Office Hadley Centre
521 Climate Program (GA01101); the Natural Environment Research Council Dtg
522 (NE/J500033/1); and the European Commission's 7th Framework Program Grant
523 Agreement 282672 (EMBRACE). The data used in this paper is available from the
524 Met Office Hadley Centre upon request. Thank you to the anonymous reviewers,
525 who improved the paper with their comments.

526

527

528 References

529

530 Anav, A., P. Friedlingstein, M. Kidston, L. Bopp, P. Ciais, P. Cox, C. Jones, M. Jung,
531 R. Myneni, and Z. Zhu (2013), Evaluating the Land and Ocean Components
532 of the Global Carbon Cycle in the CMIP5 Earth System Models, *J. Clim.*,
533 26(18), 6801–6843, doi:10.1175/JCLI-D-12-00417.1.

534 es-doc v0.9.0.1 CMIP5 Model Component Properties,
535 http://prod.static.esdoc.webfactional.com/js_client/demo/prod/comparator.htm
536 |

537 Bekessy, S. A., B. A. Wintle, D. B. Lindenmayer, M. A. McCarthy, M. Colyvan, M. A.
538 Burgman, and H. P. Possingham (2010), The biodiversity bank cannot be a
539 lending bank, *Conserv. Lett.*, 3(3), 151–158, doi:10.1111/j.1755-
540 263X.2010.00110.x.

541 Betts, R. A., P. M. Cox, M. Collins, P. P. Harris, C. Huntingford, and C. D. Jones
542 (2004), The role of ecosystem-atmosphere interactions in simulated
543 Amazonian precipitation decrease and forest dieback under global climate
544 warming, *Theor. Appl. Climatol.*, 78(1), 157–175, doi:10.1007/s00704-004-
545 0050-y.

546 Betts, R. A., N. Golding, P. Gonzalez, J. Gornall, R. Kahana, G. Kay, L. Mitchell, and
547 A. Wiltshire (2013), Climate and land use change impacts on global terrestrial
548 ecosystems, fire, and river flows in the HadGEM2-ES Earth System Model
549 using the Representative Concentration Pathways, *Biogeosciences Discuss*,
550 10(4), 6171–6223, doi:10.5194/bgd-10-6171-2013.

551 Booth, B. B. B., C. D. Jones, M. Collins, I. J. Totterdell, P. M. Cox, S. Sitch, C.
552 Huntingford, R. A. Betts, G. R. Harris, and J. Lloyd (2012), High sensitivity of
553 future global warming to land carbon cycle processes, *Environ. Res. Lett.*,
554 7(2), 024002, doi:10.1088/1748-9326/7/2/024002.

555 Brovkin, V. et al. (2013), Effect of Anthropogenic Land-Use and Land-Cover
556 Changes on Climate and Land Carbon Storage in CMIP5 Projections for the
557 Twenty-First Century, *J. Clim.*, 26(18), 6859–6881, doi:10.1175/JCLI-D-12-
558 00623.1.

559 Canadell, J. G., C. L. Quéré, M. R. Raupach, C. B. Field, E. T. Buitenhuis, P. Ciais,
560 T. J. Conway, N. P. Gillett, R. A. Houghton, and G. Marland (2007),
561 Contributions to accelerating atmospheric CO₂ growth from economic
562 activity, carbon intensity, and efficiency of natural sinks, *Proc. Natl. Acad. Sci.*,
563 104(47), 18866–18870, doi:10.1073/pnas.0702737104.

564 Ciais, P. et al. (2013), Carbon and Other Biogeochemical Cycles., in *Climate Change*
565 *2013: The Physical Science Basis. Contribution of Working Group I to the*
566 *Fifth Assessment Report of the Intergovernmental Panel on Climate Change*,
567 Cambridge University Press, Cambridge, United Kingdom.

568 Clark, D. B. et al. (2011), The Joint UK Land Environment Simulator (JULES), model
569 description – Part 2: Carbon fluxes and vegetation dynamics, *Geosci. Model*
570 *Dev.*, 4(3), 701–722, doi:10.5194/gmd-4-701-2011.

571 Collins, W. J. et al. (2011), Development and evaluation of an Earth-System model –
572 HadGEM2, *Geosci. Model Dev.*, 4(4), 1051–1075, doi:10.5194/gmd-4-1051-
573 2011.

574 Colwell, R. K., G. Brehm, C. L. Cardelús, A. C. Gilman, and J. T. Longino (2008),
575 Global Warming, Elevational Range Shifts, and Lowland Biotic Attrition in the
576 Wet Tropics, *Science*, 322(5899), 258–261, doi:10.1126/science.1162547.

577 Cox, P. ., C. Huntingford, and R. . Harding (1998), A canopy conductance and
578 photosynthesis model for use in a GCM land surface scheme, *J. Hydrol.*,
579 212–213, 79–94, doi:10.1016/S0022-1694(98)00203-0.

580 Cox, P. M. (2001), Description of the TRIFFID dynamic global vegetation model,
581 *Hadley Cent. Tech. Note*, 24, 1–16.

582 Cox, P. M., R. A. Betts, C. D. Jones, S. A. Spall, and I. J. Totterdell (2000),
583 Acceleration of global warming due to carbon-cycle feedbacks in a coupled
584 climate model, *Nature*, 408(6809), 184–187, doi:10.1038/35041539.

585 Cox, P. M., R. A. Betts, M. Collins, P. Harris, C. Huntingford, and C. D. Jones (2003),
586 Amazon dieback under climate-carbon cycle projections for the 21st century,
587 *Hadley Cent. Tech. Note*, 42.

588 Cox, P. M., R. A. Betts, M. Collins, P. P. Harris, C. Huntingford, and C. D. Jones
589 (2004), Amazonian forest dieback under climate-carbon cycle projections for
590 the 21st century, *Theor. Appl. Climatol.*, 78(1), 137–156.

591 Davies-Barnard, T., P. J. Valdes, J. S. Singarayer, and C. D. Jones (2014a), Climatic
592 impacts of land-use change due to crop yield increases and a universal
593 carbon tax from a scenario model, *J. Clim.*, 27(4), 1413–1424,
594 doi:10.1175/JCLI-D-13-00154.1.

595 Davies-Barnard, T., P. J. Valdes, J. S. Singarayer, F. M. Pacifico, and C. D. Jones
596 (2014b), Full effects of land use change in the representative concentration
597 pathways, *Environ. Res. Lett.*, 9(11), 114014, doi:10.1088/1748-
598 9326/9/11/114014.

599 Emanuel, W. R., H. H. Shugart, and M. P. Stevenson (1985), Climatic change and
600 the broad-scale distribution of terrestrial ecosystem complexes, *Clim.*
601 *Change*, 7(1), 29–43, doi:10.1007/BF00139439.

602 Essery, R., M. Best, and P. Cox (2001), *MOSES 2.2 technical documentation*,
603 Hadley Centre Technical Note.

604 Essery, R. L. H., M. J. Best, R. A. Betts, P. M. Cox, and C. M. Taylor (2003), Explicit
605 Representation of Subgrid Heterogeneity in a GCM Land Surface Scheme, *J.*
606 *Hydrometeorol.*, 4(3), 530–543, doi:10.1175/1525-7541(2003)004.

607 Gibson, L. et al. (2011), Primary forests are irreplaceable for sustaining tropical
608 biodiversity, *Nature*, 478(7369), 378–381, doi:10.1038/nature10425.

609 Good, P., C. Jones, J. Lowe, R. Betts, and N. Gedney (2012), Comparing Tropical
610 Forest Projections from Two Generations of Hadley Centre Earth System
611 Models, HadGEM2-ES and HadCM3LC, *J. Clim.*, 26(2), 495–511,
612 doi:10.1175/JCLI-D-11-00366.1.

613 Gruber, N., and J. N. Galloway (2008), An Earth-system perspective of the global
614 nitrogen cycle, *Nature*, 451(7176), 293–296, doi:10.1038/nature06592.

- 615 Hassan, R., U. R. Scholes, and N. Ash (Eds.) (2005), *Ecosystems and Human Well-Being: Findings of the Condition and Trends Working Group v. 1: Current*
616 *State and Trends*, Island Press, Washington, DC.
617
- 618 Huntingford, C., R. A. Fisher, L. Mercado, B. B. B. Booth, S. Sitch, P. P. Harris, P. M.
619 Cox, C. D. Jones, R. A. Betts, and Y. Malhi (2008), Towards quantifying
620 uncertainty in predictions of Amazon “dieback,” *Philos. Trans. R. Soc. B Biol.*
621 *Sci.*, 363(1498), 1857–1864.
- 622 Hurtt, G. et al. (2011), Harmonization of land-use scenarios for the period 1500–
623 2100: 600 years of global gridded annual land-use transitions, wood harvest,
624 and resulting secondary lands, *Clim. Change*, 109(1), 117–161,
625 doi:10.1007/s10584-011-0153-2.
- 626 Jain, A. K., P. Meiyappan, Y. Song, and J. I. House (2013), CO2 emissions from
627 land-use change affected more by nitrogen cycle, than by the choice of land-
628 cover data, *Glob. Change Biol.*, 19(9), 2893–2906, doi:10.1111/gcb.12207.
- 629 Jones, A. D. et al. (2012), Greenhouse Gas Policy Influences Climate via Direct
630 Effects of Land-Use Change, *J. Clim.*, 26(11), 3657–3670, doi:10.1175/JCLI-
631 D-12-00377.1.
- 632 Jones, C., and P. Falloon (2009), Sources of uncertainty in global modelling of future
633 soil organic carbon storage, in *Uncertainties in Environmental Modelling and*
634 *Consequences for Policy Making*, edited by P. C. Baveye, M. Laba, and J.
635 Mysiak, pp. 283–315, Springer Netherlands.
- 636 Jones, C., J. Lowe, S. Liddicoat, and R. Betts (2009), Committed terrestrial
637 ecosystem changes due to climate change, *Nat. Geosci.*, 2(7), 484–487,
638 doi:10.1038/ngeo555.
- 639 Jones, C., S. Liddicoat, and J. Lowe (2010), Role of terrestrial ecosystems in
640 determining CO2 stabilization and recovery behaviour, *Tellus B*, 62(5), 682–
641 699, doi:10.1111/j.1600-0889.2010.00490.x.
- 642 Jones, C. D. et al. (2011), The HadGEM2-ES implementation of CMIP5 centennial
643 simulations, *Geosci. Model Dev.*, 4(3), 543–570, doi:10.5194/gmd-4-543-
644 2011.
- 645 Kato, E., and Y. Yamagata (2014), BECCS capability of dedicated bioenergy crops
646 under a future land-use scenario targeting net negative carbon emissions,
647 *Earths Future*, 2014EF000249, doi:10.1002/2014EF000249.
- 648 Kurz, W. A., C. C. Dymond, G. Stinson, G. J. Rampley, E. T. Neilson, A. L. Carroll, T.
649 Ebata, and L. Safranyik (2008), Mountain pine beetle and forest carbon
650 feedback to climate change, *Nature*, 452(7190), 987–990,
651 doi:10.1038/nature06777.
- 652 Lawrence, P. J. et al. (2012), Simulating the Biogeochemical and Biogeophysical
653 Impacts of Transient Land Cover Change and Wood Harvest in the
654 Community Climate System Model (CCSM4) from 1850 to 2100, *J. Clim.*,
655 25(9), 3071–3095, doi:10.1175/JCLI-D-11-00256.1.
- 656 Liddicoat, S., C. Jones, and E. Robertson (2013), CO2 Emissions Determined by
657 HadGEM2-ES to be Compatible with the Representative Concentration

658 Pathway Scenarios and Their Extensions, *J. Clim.*, 26(13), 4381–4397,
659 doi:10.1175/JCLI-D-12-00569.1.

660 MacDonald, G. M., K. V. Kremenetski, and D. W. Beilman (2008), Climate change
661 and the northern Russian treeline zone, *Philos. Trans. R. Soc. B Biol. Sci.*,
662 363(1501), 2283–2299, doi:10.1098/rstb.2007.2200.

663 Malhi, Y., L. E. O. C. Aragão, D. Galbraith, C. Huntingford, R. Fisher, P. Zelazowski,
664 S. Sitch, C. McSweeney, and P. Meir (2009), Exploring the likelihood and
665 mechanism of a climate-change-induced dieback of the Amazon rainforest,
666 *Proc. Natl. Acad. Sci.*, 106(49), 20610–20615,
667 doi:10.1073/pnas.0804619106.

668 Maron, M., R. J. Hobbs, A. Moilanen, J. W. Matthews, K. Christie, T. A. Gardner, D.
669 A. Keith, D. B. Lindenmayer, and C. A. McAlpine (2012), Faustian bargains?
670 Restoration realities in the context of biodiversity offset policies, *Biol.*
671 *Conserv.*, 155, 141–148, doi:10.1016/j.biocon.2012.06.003.

672 Martin, G. M., M. A. Ringer, V. D. Pope, A. Jones, C. Dearden, and T. J. Hinton
673 (2006), The Physical Properties of the Atmosphere in the New Hadley Centre
674 Global Environmental Model (HadGEM1). Part I: Model Description and
675 Global Climatology, *J. Clim.*, 19(7), 1274–1301, doi:10.1175/JCLI3636.1.

676 Martin, G. M. et al. (2011), The HadGEM2 family of Met Office Unified Model climate
677 configurations, *Geosci. Model Dev.*, 4(3), 723–757, doi:10.5194/gmd-4-723-
678 2011.

679 Meinshausen, M. et al. (2011), The RCP greenhouse gas concentrations and their
680 extensions from 1765 to 2300, *Clim. Change*, 109(1), 213–241,
681 doi:10.1007/s10584-011-0156-z.

682 Moilanen, A., A. J. A. Van Teeffelen, Y. Ben-Haim, and S. Ferrier (2009), How Much
683 Compensation is Enough? A Framework for Incorporating Uncertainty and
684 Time Discounting When Calculating Offset Ratios for Impacted Habitat,
685 *Restor. Ecol.*, 17(4), 470–478, doi:10.1111/j.1526-100X.2008.00382.x.

686 Myers, N., R. A. Mittermeier, C. G. Mittermeier, G. A. B. da Fonseca, and J. Kent
687 (2000), Biodiversity hotspots for conservation priorities, *Nature*, 403(6772),
688 853–858, doi:10.1038/35002501.

689 Nishina, K. et al. (2014), Quantifying uncertainties in soil carbon responses to
690 changes in global mean temperature and precipitation, *Earth Syst Dynam.*,
691 5(1), 197–209, doi:10.5194/esd-5-197-2014.

692 De Noblet-Ducoudré, N. et al. (2012), Determining Robust Impacts of Land-Use-
693 Induced Land Cover Changes on Surface Climate over North America and
694 Eurasia: Results from the First Set of LUCID Experiments, *J. Clim.*, 25(9),
695 3261–3281, doi:10.1175/JCLI-D-11-00338.1.

696 Pongratz, J., C. H. Reick, R. A. Houghton, and J. I. House (2014), Terminology as a
697 key uncertainty in net land use and land cover change carbon flux estimates,
698 *Earth Syst Dynam.*, 5(1), 177–195, doi:10.5194/esd-5-177-2014.

- 699 Prentice, I. C., M. T. Sykes, and W. Cramer (1991), The Possible Dynamic Response
700 of Northern Forests to Global Warming, *Glob. Ecol. Biogeogr. Lett.*, 1(5),
701 129–135, doi:10.2307/2997426.
- 702 Le Quéré, C. et al. (2014a), Global carbon budget 2013, *Earth Syst. Sci. Data*, 6(1),
703 235–263, doi:10.5194/essd-6-235-2014.
- 704 Le Quéré, C. et al. (2014b), Global carbon budget 2014, *Earth Syst. Sci. Data*
705 *Discuss.*, 7(2), 521–610, doi:10.5194/essdd-7-521-2014.
- 706 Reid, C. T. (2013), Between Priceless and Worthless: Challenges in Using Market
707 Mechanisms for Conserving Biodiversity, *Transnatl. Environ. Law*, 2(02),
708 217–233, doi:10.1017/S2047102512000210.
- 709 Shevliakova, E., S. W. Pacala, S. Malyshev, G. C. Hurtt, P. C. D. Milly, J. P.
710 Caspersen, L. T. Sentman, J. P. Fisk, C. Wirth, and C. Crevoisier (2009),
711 Carbon cycling under 300 years of land use change: Importance of the
712 secondary vegetation sink, *Glob. Biogeochem. Cycles*, 23(2),
713 doi:10.1029/2007GB003176.
- 714 Soja, A. J., N. M. Tchepakova, N. H. F. French, M. D. Flannigan, H. H. Shugart, B. J.
715 Stocks, A. I. Sukhinin, E. I. Parfenova, F. S. Chapin III, and P. W. Stackhouse
716 Jr. (2007), Climate-induced boreal forest change: Predictions versus current
717 observations, *Glob. Planet. Change*, 56(3–4), 274–296,
718 doi:10.1016/j.gloplacha.2006.07.028.
- 719 Stocker, T. F., D. Qin, M. Plattner, J. Boschung, A. Nauels, S. K. Allen, M. Tignor, V.
720 Bex, P. M. Midgley, and Y. Xia (Eds.) (2013), *IPCC, 2013: Climate Change*
721 *2013: The Physical Science Basis. Contribution of Working Group I to the*
722 *Fifth Assessment Report of the Intergovernmental Panel on Climate Change*,
723 Cambridge University Press, Cambridge, United Kingdom.
- 724 Taylor, K. E., R. J. Stouffer, and G. A. Meehl (2012), An Overview of CMIP5 and the
725 Experiment Design, *Bull. Am. Meteorol. Soc.*, 93(4), 485–498,
726 doi:10.1175/BAMS-D-11-00094.1.
- 727 Thomson, A. M., K. V. Calvin, L. P. Chini, G. Hurtt, J. A. Edmonds, B. Bond-
728 Lamberty, S. Froking, M. A. Wise, and A. C. Janetos (2010), Climate
729 mitigation and the future of tropical landscapes, *Proc. Natl. Acad. Sci.*,
730 107(46), 19633–19638, doi:10.1073/pnas.0910467107.
- 731 Thomson, A. M. et al. (2011), RCP4.5: a pathway for stabilization of radiative forcing
732 by 2100, *Clim. Change*, 109(1–2), 77–94, doi:10.1007/s10584-011-0151-4.
- 733 Van Vuuren, D. et al. (2011), The representative concentration pathways: an
734 overview, *Clim. Change*, 109(1), 5–31, doi:10.1007/s10584-011-0148-z.
- 735 Walther, G.-R., E. Post, P. Convey, A. Menzel, C. Parmesan, T. J. C. Beebee, J.-M.
736 Fromentin, O. Hoegh-Guldberg, and F. Bairlein (2002), Ecological responses
737 to recent climate change, *Nature*, 416(6879), 389–395, doi:10.1038/416389a.
- 738 Wilkenskjeld, S., S. Kloster, J. Pongratz, T. Raddatz, and C. Reick (2014),
739 Comparing the influence of net and gross anthropogenic land use and land
740 cover changes on the carbon cycle in the MPI-ESM, *Biogeosciences Discuss*,
741 11(4), 5443–5469, doi:10.5194/bgd-11-5443-2014.

Woodward, F. I., M. R. Lomas, and R. A. Betts (1998), Vegetation-climate feedbacks in a greenhouse world, *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, 353(1365), 29–39, doi:10.1098/rstb.1998.0188.

Figure Captions

Figure 1. Conceptual diagram of the simulations and how the different diagnostics used in the paper are calculated.

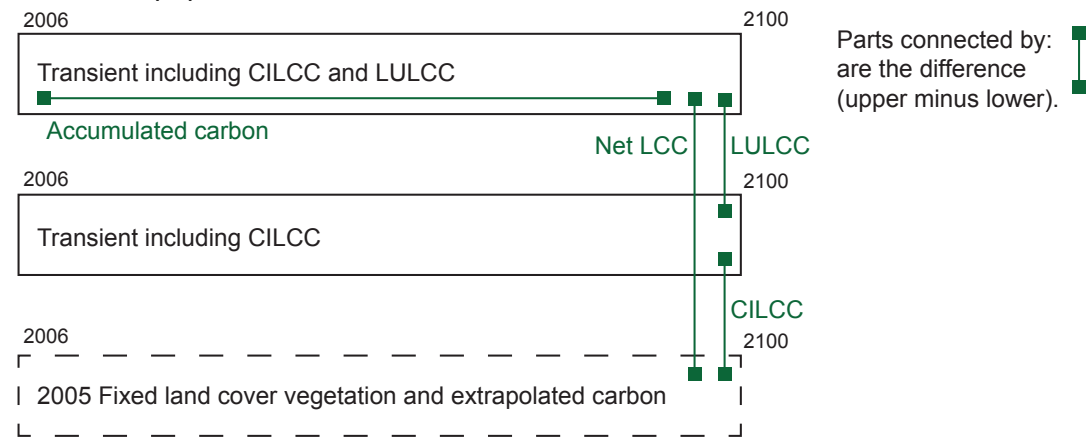


Figure 2. Changes in forest fraction (in per cent of total global land area) (top) globally, (middle) temperate/boreal forest area (33.75 N – 83.75 N, mid to high latitude Northern hemisphere) and (bottom) the Tropics (16.25 S – 21.25 N). For left column (a – c) 2050-2006 and for right column (d – f) 2100 – 2006.

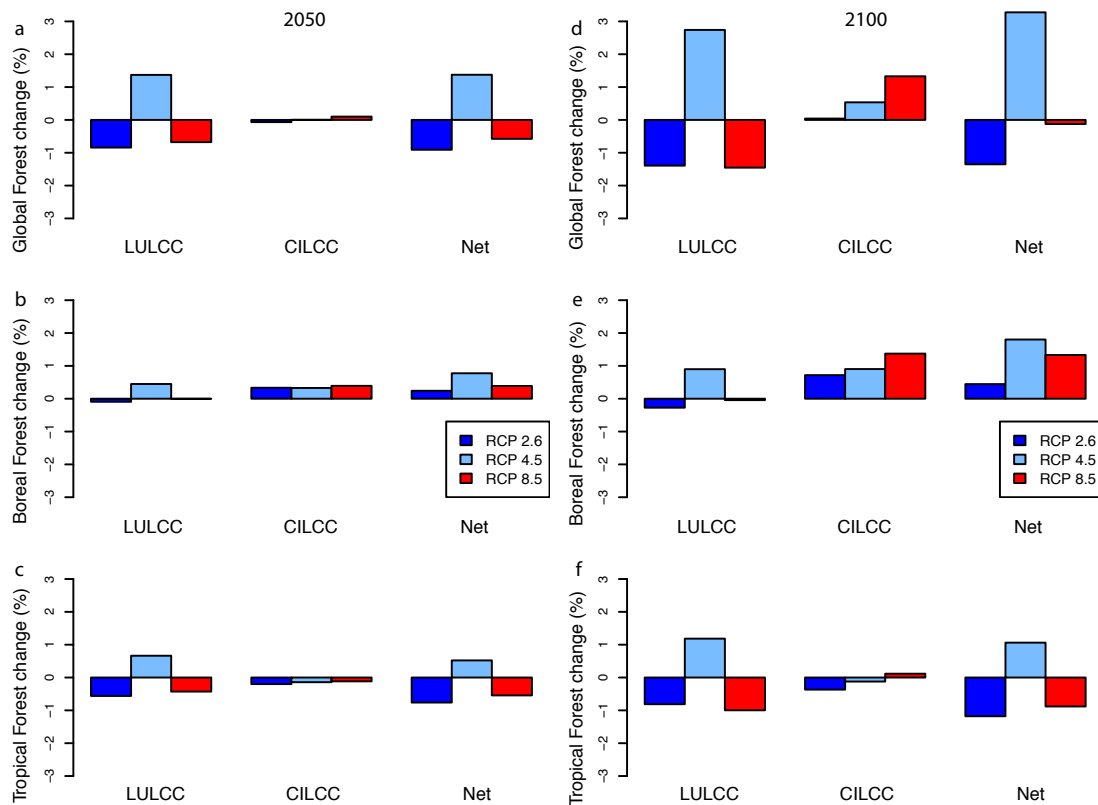


Figure 3. The LULCC 2005 to 2100, encompassing the agricultural fraction changes (crop and pasture land).

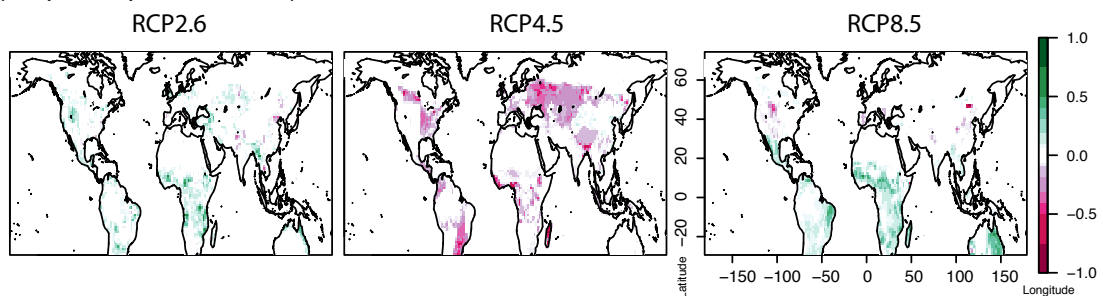


Figure 4. Change in woody veg surface types, 2100 – 2005 from CILCC. Rows from the top: Broadleaf trees, needleleaf trees, shrubs.

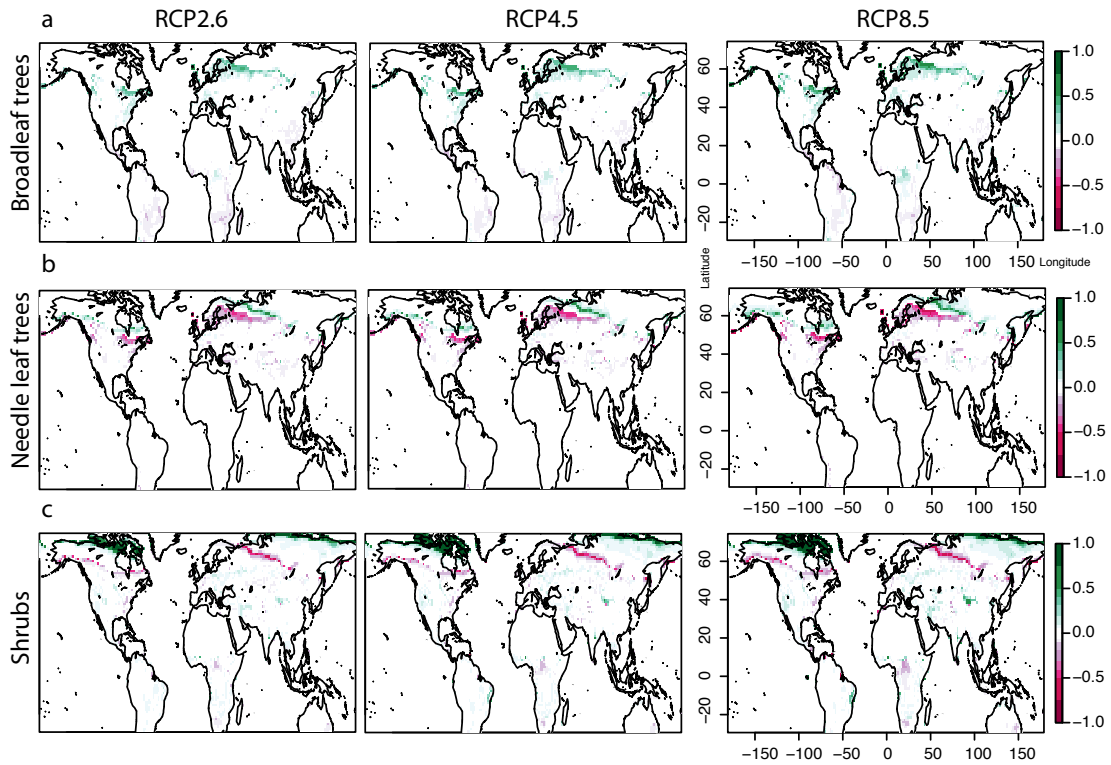


Figure 5. Change in selected non woody vegetation/surface types from CILCC, 2100 – 2005. Rows from the top: C_3 grasses, C_4 grasses, bare soil.

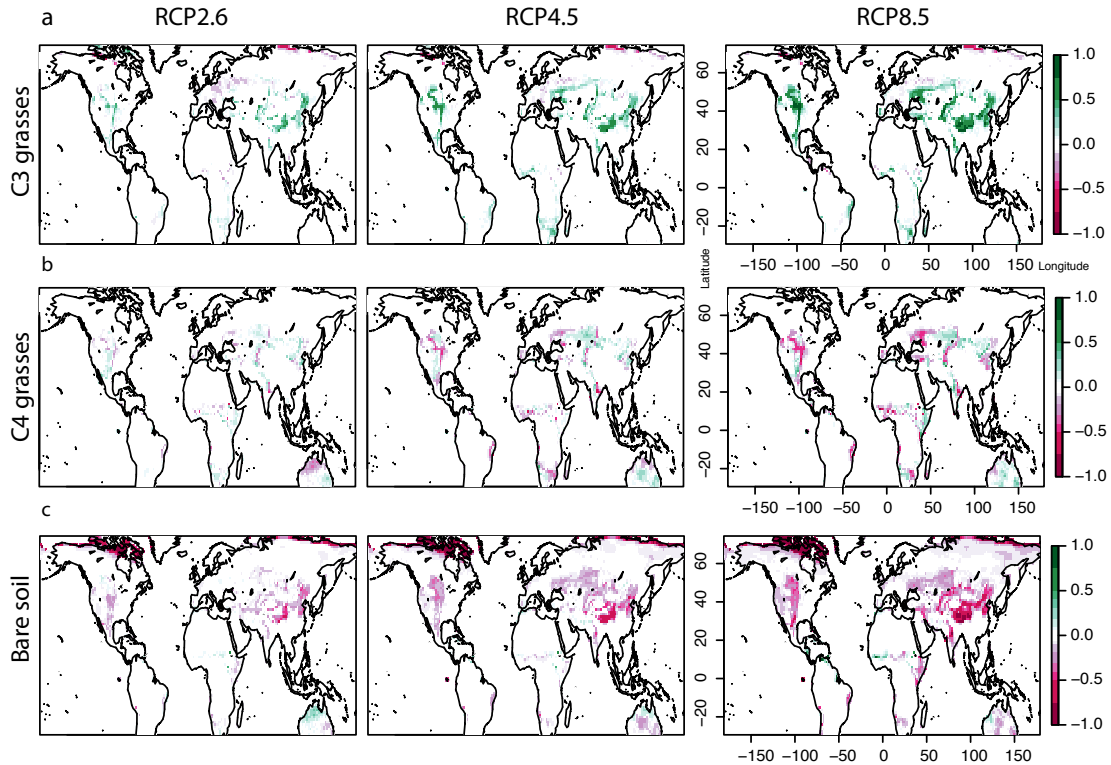
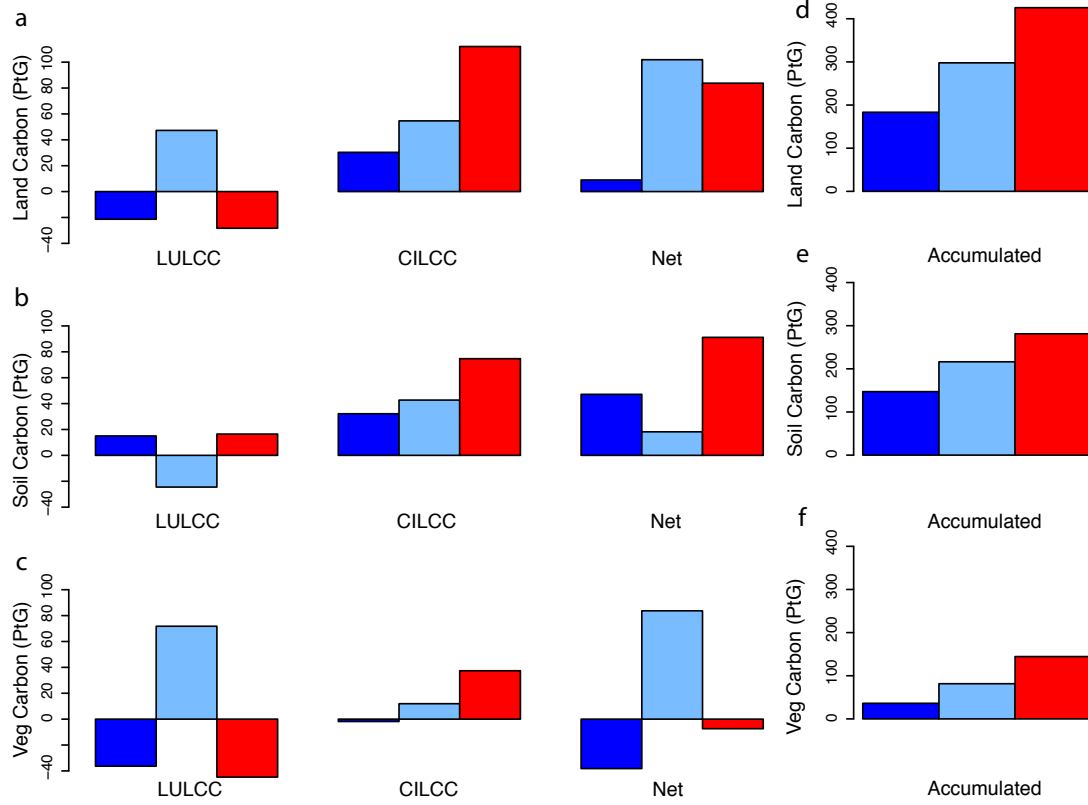


Figure 6. Anomaly of total global land carbon storage changes from different sources, 2100 - 2005. For a) – c) LULCC, CILCC and Net (LULCC+CILCC). For d) – f) the Accumulated carbon storage change (from all land surface, not just LCC). Separated into: a) and d) vegetation and soil carbon; b) and e) soil carbon; c) and f) vegetation carbon. Note that the scale for d) to f) is 4 times larger than for a) – c).



777

778